

**Australia's National Science Agency** 

# **Application to release the** cabomba weevil Hydrotimetes natans for the biological control of Cabomba caroliniana in Australia

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## Executive summary

*Cabomba caroliniana* (Cabombaceae) is a submerged aquatic weed of permanent and slowmoving freshwater bodies. This weed affects water quality, recreational activities and public safety, and is a Weed of National Significance. It has a wide distribution in Australia from Melbourne to Darwin and can grow in a range of climate conditions from monsoonal tropics to cold temperate environments.

The existing control options for *C. caroliniana* are either unfeasible or expensive. For instance, herbicides are highly restricted around public/potable water supplies and mechanical removal using machines is generally prohibitively expensive. The Australian Weed Committee (now subsumed under the Environment and Invasives Committee) endorsed *C. caroliniana* as a target for biological control in Australia in 2005. Biological control (biocontrol) may be the only viable option for management of *C. caroliniana* given the difficulty with existing control options and their safety to use in water supplying dams.

The cabomba weevil, *Hydrotimetes natans* has been identified as a potential biocontrol agent for *C. caroliniana*. We imported *H. natans* from Paraguay and Argentina under quarantine conditions in Australia to conduct comprehensive host-specificity testing on a broad range of plant species, including native Australian species, selected based on their phylogenetic relationships to the target weed. A total of 17 plant species from the families Cabombaceae, Nymphaeaceae and Hydatellaceae were tested − 14 in the laboratory in Australia and 3 in the laboratory in Argentina. In addition, field host-specificity assessments were performed at four sites in Argentina and Paraguay where the weevil was recorded on *C. caroliniana*. Co-occurring non-target aquatic species with *C. caroliniana* (e.g. *Egeria najas, Nymphoides indica, Nymphaea prolifera, Salvinia minima* and *Ludwigia grandiflora*) were examined for the presence of *H. natans* and any sign of feeding by larvae and adults.

Results from field observations and laboratory trials are briefly outlined below:

Field host-specificity: In the field in Argentina and Paraguay, observations revealed the presence of *H. natans* almost exclusively on *C. caroliniana* except for a single *H. natans* adult observed on *N. prolifera* adjacent to *C. carolinian*a. However, no feeding on *N. prolifera* was noticed which suggested that it was likely a casual occurrence.

Adult and larval feeding on leaf discs/sprigs: Feeding lesions caused by adult *H. natans* were observed on *C. caroliniana, Brasenia schreberi, Nymphaea caerulea, N. gigantea*, *N. nouchali, N. prolifera and Victoria cruziana* but not on *N. mexicana and C. caroliniana* var *flavida*. Most of the feeding lesions on non-target plants were superficial and exploratory. Larval feeding trials on *C. caroliniana var. flavida, Nymphaea prolifera, N. caerulea* and *V. cruziana* demonstrated larvae are highly specific and unable to feed on non-target species. All larvae on non-target species died within four to five days of exposure to these species.

No-choice trials: Larval feeding, oviposition and larval development to adult occurred consistently on *C. caroliniana*. No oviposition occurred on any of the *Nymphaea* or *Trithuria* test plant species and hence no progeny development was observed. In *B. schreberi*, oviposition occurred on four of the six replicates tested. Among the four replicates that showed evidence of oviposition, larval feeding was noticed on three replicates, and pupation and adult emergence was observed in only one replicate.

Choice and continuation trials: Choice trials with *C. caroliniana* and *B. schreberi* suggested partial lifecycle development of *H. natans* on *B. schreberi*. Oviposition and larval development were observed on two of the five replicates tested. However, larval development to pupation was observed on only one replicate of *B. schreberi* with two pupae recorded. Only one of the two pupae metamorphosed into an adult, which however died soon after emergence, and the other pupa did not emerge as adult. In continuation trials, despite its exposure for 150 days (a duration equivalent to three generations of the weevil on *C. caroliniana*), *B. schreberi* did not sustain a *H. natans* population. Only one pupa was observed on one replicate of *B. schreberi* despite egg laying observed in three replicates. In contrast, a healthy and reproducing colony of *H. natans* was maintained on *C. caroliniana*, which has yielded five generations in the eight-month period between April 2019 to December 2019, the same period over which the laboratory testing was undertaken.

In summary, results from a suite of laboratory-based host-specificity testing in the native range and in a quarantine glasshouse and laboratory in Australia, as well as field observations in the native range demonstrated that *H. natans* has a high degree of specificity towards the target weed *C. caroliniana*. Only the native species, *B. schreberi* supported partial development of *H. natans*, with the insect completing its lifecycle in only one replicate and F1 adults neither surviving nor developing through to the F2 generation. In a subsequent continuation trial with this species, *H. natans* could not sustain its population beyond the F1 generation of offspring. Further, the feeding damage caused by *H. natans* (adults and larvae) on *B. schreberi* was minimal compared to that occurring on *C. caroliniana*. Based on the series of trials undertaken in this study, we conclude that the level of risk *H. natans* poses to non-target native and introduced species in Australia is negligible and that *H. natans* will potentially be an effective biological control agent for *C. caroliniana*.

The decision tree presented below shows the rationale behind the series of trials undertaken in this study and summarizes the level of risk likely to occur for each test plant species.



Decision tree used to determine the types of host-specificity tests to be undertaken for *Hydrotimetes natans* for the target species, *Cabomba caroliniana*. The types of tests carried out, the outcome of host plant testing on each non-target plant species and conclusions made based on outcome is shown above

## 1 Information on the target species, *Cabomba caroliniana*

#### 1.1 Taxonomy

#### **1.1.1 Botanical name**

*Cabomba caroliniana* A. Gray

#### **1.1.2 Common name**

The plant is usually referred to as cabomba or water fanwort in Australia, as Carolina watershield, water fanwort, Washington grass or fish grass in the USA, and cabomba or water nettle in South America.

#### **1.1.3 Relationships**

*Cabomba caroliniana* is a member of the Order Nymphaeales, a basal clade of angiosperms, which is distantly related to the rest of the angiosperm clade (Saarela et al. 2007; Iles et al. 2014). There are only two species in the Cabombaceae family in Australia; one being *C. caroliniana*, the other in another genus *Brasenia schreberi* J.F. Gmel (watershield). Other families in the Nymphaeales are Nymphaeaceae (the water lilies) which is comprised of 23 species of *Nymphaea* L. and the Hydatellaceae which is comprised of 10 species of its only genus *Trithuria* Hook.f. in Australia (Jacobs & Porter 2007; Löhne et al. 2008; Sokoloff et al. 2011) (Figure 1).



**Figure 1 Phylogeny indicating the relationships with the order Nymphaeales and its relationships to other angiosperm groups (Stevens 2001 onwards; Chase et al. 2016)**

#### **1.1.4 Close relatives native in Australia**

Cabombaceae contains only two genera in Australia, *Cabomba* Aubl. (the genus of the target weed, *C. caroliniana*) and *Brasenia* Schreb (Appendix I). *Brasenia* is a monotypic genus with only one species *B. schreberi*. *Brasenia schreberi* is distributed across all continents (eastern Asia, Australia, Africa, the West Indies, and South, Central, and North America) except Europe (extinct) and Antarctica (Drzymulska 2018). It occurs in shallow lakes and ponds, as well as slow moving streams and has a range extending from tropical to nearly boreal habitats but is typically more abundant at higher latitudes (Drzymulska 2018). It is considered native to north America, central America, Africa, Asia and Australia (Figure 2). In Australia, *B. schreberi* has a scattered/localized distribution in Queensland (QLD), VIC (Victoria) and NSW (New South Wales), tending to occur in greater abundance in temperate water bodies (Orgaard 1991).



**Figure 2 (A) Native range of** *Brasenia schreberi* **(Source: Royal Botanic Gardens, Kew; http://plantsoftheworld.online/taxon/urn:lsid:ipni.org:names:605270-1#distribution-map); (B) Distribution of** *B.*  *schreberi* **in Australia (source: Atlas of Living Australia); this species appears to be more abundant in temperate and subtemperate parts of its range.**

Within the family Nymphaeaceae, the only genus present in Australia is *Nymphaea* (Appendix I). Plants of this genus are known commonly as water lilies, and are present in temperate, subtropical, and tropical regions in all continents except Antarctica (Heslop-Harrison 1955). Numerous hybrids and cultivars have been developed and are popular as ornamentals in water gardens. In Australia, several species (e.g. *N. mexicana* Zucc.*, N. caerulea* Sav.*, N. alba* L.) are regarded as environmental weeds (Hussey et al. 2007).

The Hydatellaceae family is only represented by the genus *Trithuria*. *Trithuria* are diminutive, moss-like, subaquatic plants; most of the *Trithuria* are native to Australia, but some are native to India (*T. konkanensis* S.R. Yadav & Janarth) and New Zealand (*T. inconspicua* Cheeseman). In Australia, *Trithuria* is found in Western Australia (WA), Northern Territory (NT), QLD, South Australia (SA), VIC, some parts of NSW and Tasmania (TAS) (Atlas of Living Australia, 2020a). Some *Trithuria* species are endemic to certain localities and listed as endangered or threatened species (Appendix I). For example, *T. submersa* Hook.f. occurs in WA, SA, VIC and NSW, and is listed as a threatened species. In Tasmania, *T. submersa* occurs in marshy habitat in the Midlands, Central Highlands and the north-east of the State. *Trithuria occidentalis* Benth. is endemic and strictly confined to Ellenbrook and possibly to Upper Swan in WA and is ranked as a critically endangered species.

### 1.2 Description

*Cabomba caroliniana* is a perennial submerged aquatic macrophyte. Leaves are oppositely arranged and are finely divided, forming feathery, fan-shaped structures (Figure 3A). It has small, elongate floating leaves at distal ends of the plant usually borne on flowering branches (Figure 3B). Slender, round or slightly compressed stems are usually 2–4 mm in diameter and about 1–2 m long but may be up to 10 m (Figure 3C). The young stems are usually pubescent with short white or rust-coloured hairs. Bisexual and protogynous (having female reproductive organs come to maturity before the male) long-stemmed flowers are white in colour and borne just above the water surface on pubescent stalks (Figure 3D). The sepals and petals are about 1.25 cm across. The petals are auriculate at their bases, and obovate in shape. Roots arise from the rhizome and adventitious roots are often produced at lower stem nodes (Wilson et al. 2007).



**Figure 3** *Cabomba caroliniana***: (A) Oppositely arranged feathery fan-shaped leaves, (B) elongate floating leaves on distal end of the plant, (C) thin and long stems submerged, and (D) white flower with yellow centre (Source: CSIRO).**

## 1.3 Distribution

#### **1.3.1 Native range**

*Cabomba caroliniana* is native to South America occurring in southern Brazil, Uruguay, Paraguay and north-eastern Argentina at the Parana river floodplain in South America (Figure 4). Occurrence in the southeast coast of Brazil is considered as part of the native range (Ørgaard 1991). *Cabomba caroliniana* was considered native to the south-eastern United States, however, the disjunct nature of this population from the native range population suggests that it was perhaps introduced and subsequently naturalized in the United States (Mackey & Swarbrick 1997).



**Figure 4 Global distribution of** *Cabomba caroliniana* **(Source: https://www.gbif.org/occurrence/map?taxon\_key=2882443)**

#### **1.3.2 Australian range**

The current distribution of *C. caroliniana* in Australia extends along the east coast from the Atherton Tablelands to Melbourne, and there are localized infestations near Darwin (Figure 5A). The potential distribution in Australia is much greater than the current distribution. Most of coastal Australia, except for the north-west, is predicted to be excellent habitat for *C. caroliniana* based on CLIMEX model (Mackey 1996). Recently, we developed a model using an Ecoclimatic Index (based on minimum and maximum temperature) in CLIMEX (Kriticos et al. 2015), which predicted that much of coastal Australia would be highly suitable for *C. caroliniana* establishment (Figure 5B); this is similar to previous CLIMEX predictions. The Ecoclimatic Index map shows the suitability of Australian regions where permanent water bodies are present. This map includes all perennial water bodies, including those of marginal suitability in terms of water quality for *C. caroliniana*, and should be viewed as an outer limit of projected distribution of *C. caroliniana* (Figure 5B).

In Australia, most *C. caroliniana* infestations occur in southern QLD and the northern NSW hinterland. In QLD it occurs in shallow, permanently flowing creeks and deep, slow-flowing pools of coastal river systems. The largest and most dense infestations occur in Lake MacDonald and Lake Kurwongbah on the Sunshine Coast, but significant infestations also occur in northern QLD (Figure 5C & D).

*Cabomba caroliniana* is also growing in numerous creeks and river systems on the NSW North Coast and in lakes in central VIC. In NSW, *C. caroliniana* in commonly recorded in Grafton, and occasional and localised populations were recorded along the coast from Grafton to Sydney. At Lake Nagambie in VIC, the infestation has spread and covers about 50–60 ha. *Cabomba caroliniana* has also been the subject of an expensive eradication campaign at Marlow Lagoon at Palmerston in the NT. Two major infestations have been recorded in the NT. The first at Marlow Lagoon, Palmerston, was successfully eradicated in 2003. A second, and persisting, infestation was located in an isolated section of Darwin River in 2004. The eradication effort continues in the NT.

*Cabomba caroliniana* infestations have not yet been found in WA, SA, TAS or the ACT. However, based on climate and the availability of suitable freshwater bodies, it could easily spread beyond its current distribution, especially across southern and eastern Australia. It can potentially infest waterways from Cape York to Hobart and from Sydney to Melbourne, Perth and the Ord River irrigation systems. Most freshwater floodplains, billabongs and water reservoirs in the NT are potentially susceptible to *C. caroliniana* invasion (Department of Land Resource Management, 2015).



**Figure 5 (A) Current distribution of** *Cabomba caroliniana* **in Australia (source: Atlas of Living Australia), (B) Potential distribution represented as Ecoclimatic index (EI) with temperature parameters derived from the growth experiment and stress parameters derived from the global distribution of** *C. caroliniana* **(Kriticos et al. unpublished). Redder colours indicating areas of greater climatic suitability for** *C. caroliniana.* **The masked areas do not present perennial water bodies based on National Surface Water Information (https://www.ga.gov.au/scientific-topics/national-location-information/national-surface-water-information), and** 

**hence unlikely to be invaded by** *C. caroliniana***, (C) infestation in Lake MacDonald, Queensland (Source: CSIRO), (D) infestation in Siebs Dam, Cooroy, Queensland (Source: CSIRO).**

## 1.4 Ecology

*Cabomba caroliniana* grows well in standing water and can grow in slow flowing streams as well. It is often found along the margins of deeper water. It can grow at the rate of 5 cm a day under ideal conditions and can adapt to and survive under a wide range of environmental conditions (Tarver and Sanders 1977, Sanders 1979). In its native range, *C. caroliniana* prefers habitats with slow moving water and a thick substrate of organic matter. In Australia, it has been found growing in shallow reservoirs with silt substrates (Garraty et al.1996; Mackey and Swarbrick 1997; Diatloff

and Anderson 1995). *Cabomba caroliniana* is not usually found on sand or hard substrates (Sheldon 1994) and if so, grows with reduced vigour (Garraty et al. 1996; Mackey and Swarbrick 1997).

Sexual reproduction of *C. caroliniana* is not well understood and probably a secondary mode of reproduction. There is much uncertainty surrounding flowering, seed production, seed viability and germination. In north QLD *C. caroliniana* flowers continuously throughout the year but it is unknown if flowers set seeds or not. Seeds and seedlings have been found near Darwin (NT), and it was hypothesised that only the populations in Darwin (NT) reproduce by seed (Schooler et al. 2006; Dugdale et al. 2013). However, a QLD population was recently observed to set seeds in the laboratory; germination and viability of these seeds are being studied (TO Bickel pers. comm.).

## 1.5 Importance

#### **1.5.1 Beneficial aspects**

*Cabomba caroliniana* has no beneficial aspects. It was used as an aquarium plant and sold through the aquarium and nursery industry. However, it is prohibited now in all states and territories (except VIC, where it is classified as a high-risk environmental weed) because of its invasiveness.

#### **1.5.2 Detrimental aspects**

*Cabomba caroliniana* is a Weed of National Significance and regarded as one of the worst weeds in Australia (Thorpe & Lynch 2000). It forms dense monocultures in aquatic ecosystems, which reduces light availability through the water column to native species. Extensive infestations have been shown to displace submerged vegetation, including native species such as *Vallisneria nana*  R.Br., *Hydrilla verticillata* (L.f.) Royle, *Najas spp.*, *Potamogeton spp.* and *Ceratophyllum spp.* in QLD (Mackey and Swarbrick 1997). *Cabomba caroliniana* reduces the water holding capacity of dams supplying drinking water and water for agriculture. It also reduces water quality through discolouration and tainting and changes the nutrient profile, which is detrimental to aquatic fauna (Mackey and Swarbrick 1997). Dense infestations of *C. caroliniana* are associated with declines of platypus and water rat populations in northern QLD and Mary River Cod populations in Lake Macdonald, QLD (GHD, 2008). Elsewhere in Canada, the United States and China, *C. caroliniana* is suspected to negatively affect fish and invertebrate populations in addition to displacing native vegetation (Mackey & Swarbrick 1997; Wilson et al. 2007; Hogsden et al. 2007).

Recreational activities are also negatively affected by *C. caroliniana,* making swimming activities unsafe and creating a workplace health and safety issue for individuals working in and around water bodies infested by the weed (Mackey 1996; Mackey & Swarbrick 1997). The tourism potential of lakes infested with *C. caroliniana* (fishing, paddling and boating) and land values of properties surrounding infestations are greatly reduced (Mackey 1996; Mackey & Swarbrick 1997; Schooler et al. 2009).

The cost of managing *C. caroliniana* is substantial because it invades freshwater and potable water systems, where the use of chemical herbicides is restricted due to risks of non-target impacts.

## 1.6 Current control methods

#### **1.6.1 Mechanical control**

Using harvesters to clear thick mats of *C. caroliniana* can temporarily manage this weed but repeated removal is often necessary as plants grow quickly after removal. Such mechanical removal is expensive and hence only used in recreational areas to improve public access. For example, the cost of managing *C. caroliniana* using mechanical removal in infested dams in QLD is estimated to be more than \$250,000/year (SEQ Water, pers. comm). Further, mechanical removal should be practiced with caution because it can potentially spread the weed beyond the infested water body, as *C. caroliniana* can easily break into stem fragments when disturbed and recolonise the treated area or invade adjacent non-infested areas. Hand pulling by divers is also practiced in Lake Macdonald, QLD. While it is suitable method for isolated plants and small areas, regrowth in the cleared area has been noticed to occur within two weeks.

#### **1.6.2 Drying or shading**

Shading created by floating blankets made from builders' black plastic has been suggested as a control option for *C. caroliniana* (Schooler 2008). However, shading needs to be in place for a long time (3 or 4 months) and is feasible only in small infestations like farm dams. The cost of this management tactic is prohibitive for large infestations.

#### **1.6.3 Chemical control**

Several herbicides (e.g. endothall, 2,4-D, 2,3,5-T, Fenoprop, diquat) have been tested and found to kill *C. caroliniana* along with other aquatic weeds. In Ewen Maddock Dam in south-east QLD, 2,4-D n-butyl ester plus diatomaceous earth mixed at 1 part to 20 parts of water, injected 2 m below the water surface (final concentration of 10 ppm clay/2,4-D active ingredient) has been found to provide effective control of *C. caroliniana* (Diatloff and Anderson 1995). In QLD, application of 2 ppm (2 mg/L) of carfentrazone-ethyl over/into the water is also recommended for controlling *C. caroliniana*. It is noteworthy that herbicides are difficult to apply to submerged plants, retreatment is often necessary to maintain the control and herbicide use is highly restricted in potable water systems (Madsen 1996).

#### **1.6.4 Drawdown**

Lowering of the water level (drawdown) to expose stems and leaves of *C. caroliniana* is effective in drying out the weed. In combination with herbicides and shading, draw down can be effective to manage *C. caroliniana* in smaller water bodies. However, drawdown needs to be in place for several months and is not suitable during the wet season or for large water bodies (Dugdale et al. 2013).

## 1.7 Information on all other relevant Commonwealth, State, Territory legislative controls of the target species

*Cabomba caroliniana* is prohibited entry to Australia for the end use of seeds for sowing or nursery stock (BICON 2020). It is prohibited in all states and territories. In VIC, it has been declared a noxious weed under the Catchment and Land Protection Act 1994; movement and sale are prohibited anywhere in the state. In the ACT, *C. caroliniana* is a class 1 notifiable pest plant and class 4 prohibited pest plant under the Pest Plants and Animals Act 2005. It is on the potential pest plant list and is under ongoing monitoring; its importation, propagation and supply are prohibited in the ACT. In NSW, it is classified among 'priority weeds' under the Biosecurity Act 2015; restrictions on trade and movement apply to all parts of the plant including cuts, cultivars and hybrids. It must not be sold anywhere in NSW, and people that buy or sell are committing an offence under the *Biosecurity Act 2015* that carries large penalties. In QLD, *C. caroliniana* is a restricted invasive plant under the Biosecurity Act 2014. Under the general biosecurity obligation, everyone is responsible to take all reasonable and practical steps to minimise the risks associated with *C. caroliniana* in QLD and not to do anything that might make the risks worse. It is not permitted to be sold or released into the environment without a permit (BQ, 2016).

*Cabomba caroliniana* is an eradication target in all areas of NT (Class A) and is not to be introduced to any areas (Class C) under section 7 of the Weeds Management Act 2001. Eradication is targeted on small infestations where feasible. It is illegal to transport, sell, buy or propagate *C. caroliniana* plants or seeds in NT. In WA, it is classified as a Declared Pest, Prohibited -s12 under the Biosecurity and Agriculture Management Act 2007; may only be imported and kept subject to permits. Permit conditions applicable to some species may only be appropriate or available to research organisations or similarly secure institutions. Introduction or movement and supply or advertising supply are prohibited, and plants are to be eradicated from part or all of WA. In SA, *C. caroliniana* is declared in category 1 under the Natural Resources Management Act 2004. The movement or transport of the plant on a public road by itself or as a contaminant, its entry to SA, or sale by itself or as a contaminant are prohibited. Notification of infestations is necessary to ensure these are destroyed. Landowners are required to destroy any *C. caroliniana* plants growing on their properties (Declared Plant Policy 2004). In TAS, *C. caroliniana* is a Declared weed under the Weed Management Act 1999; importation, sale, distribution, movement and storage are prohibited; plants/infestations are to be reduced, eradicated or restricted. In VIC, this weed is classified as a high-risk environmental weed (White et al. 2018) and declared a noxious weed (Agriculture Victoria 2019)*.*

## 1.8 Stakeholder consultation

*Cabomba caroliniana* has been traded by aquarium nurseries previously in Australia but now it is declared as either a prohibited or noxious weed throughout Australia. Aquarium traders are aware of the legal status of the weed in their respective states and not selling *C. caroliniana*. It was clear from consultations with aquarium traders that *C. caroliniana* is not a preferred aquarium plant and hence there is no trade implications because the weed is under active management across the nation.

## 1.9 When the target species was approved for biological control

The Australian Weeds Committee (now subsumed under the Environment and Invasives Committee) endorsed *C. caroliniana* as a target for biological control in Australia in 2005 (https://weeds.ala.org.au/target.html).

## 1.10 History of biological control of *C. caroliniana*

No agents have been released in Australia for the biological control of *C. caroliniana* thus far. In 2004, CSIRO and the USDA-ARS South American Biological Control Laboratory (currently, Fundación para el Estudio de Especies Invasivas (FuEDEI)) Hurlingham, Argentina, began surveys for natural enemies of *C. caroliniana* in South America. Three candidate agents; the cabomba weevil *Hydrotimetes natans* Kolbe, and the moth species *Paracles* sp. and *Paraponyx diminutalis* (Snellen) were prioritized for further investigation. *Paracles* sp. and *P. diminutalis* were rejected because of non-target feeding on other aquatic plants in laboratory no-choice and preference tests conducted in Argentina (Schooler et al. 2012).

Preliminary no-choice feeding and oviposition tests with *H. natans* on *Egeria densa* Planch., *Ceratophyllum demersum* L. and *Myriophyllum aquaticum* (Vell.) Verdc*.* were performed under laboratory conditions in Argentina. Eggs and several feeding punctures were found on *C. caroliniana* tips during these tests, but none were found on the other species (Cabrera-Walsh et al. 2011). Species co-occurring with *C. caroliniana* in the field in the native; *E. densa, E. naias* Planch.*, Najas* sp*.* L., *Cabomba haynesii* Wiersema*, Nymphoides indica* (L.) Kuntze*, Potamogeton illinoensis*  Morong and *P. gayi* A.Benn*, C. demersum* L.*, Urticularia platensis* Speg. and *U. foliosa* L. were also surveyed for presence of *H. natans*. *Hydrotimetes natans* was recorded only on *C. caroliniana* and not on any other species (Cabrera-Walsh et al. 2011). These native range studies demonstrated that *H. natans* was host specific to *C. caroliniana*. Based on these field observations and results from preliminary laboratory tests, *H. natans* was imported into quarantine in Australia for further testing but rearing of the weevil was not successful, and the project was put on hold in 2006.

## 2 Information on the potential agent, *Hydrotimetes natans*

*Hydrotimetes natans* is native to South America and is observed to persist solely on *C. caroliniana*  in its native range (Cabrera-Walsh *et al*. 2011). Adults feed on the tips and stems of the plant, causing only limited damage to the plant. Larvae are stem-miners that can cause considerable damage to the stems of the plant and tissue necrosis. Adult weevils are dark brown dorsally with light brown-cream venters and are around 4.6mm in length. They spend the majority of their time clinging to the submerged parts of the plant, although they are observed often in nature resting or mating on the flowers.

### 2.1 Taxonomy

Order: Coleoptera

Tribe: Bagoini

Family: Curculionidae

Subfamily: Erirhininae

Genus: *Hydrotimetes*

Species: *natans*

## 2.2 Description

The Erirhininae subfamily, to which *H. natans* belongs, has the following features (Kuschel, 1971). *Aedeagus*: the aedeagal body frequently divided into a dorsal and a ventral plate by a lateral membrane (as in Orthoceri or primitive weevils). Apodemes are broad and long, usually with a bifurcation at base, whose dorsal branch frequently fused with the opposite one thus forming a dorsal bridge or arch; this bridge separates from or fused to the median dorsal portion of the aedeagal body. These features that are usually a characteristic of the Orthoceri families (and Rhynchophorinae) show that Erirhininae share most characters of the last abdominal segments and genitalia of the males with the Orthoceri families. *Tegmen*: very large, nearly as long as or longer than aedeagus (including apodemes). Apodeme always long and broad. Ring strongly proclinate in lateral aspect. Parameres often very large and fringed with abundant long hairs (as in Orthoceri), seldom somewhat reduced or even absent. *Tergite 8*: hidden under to well exposed beyond tergite 7. *Sternite 8*: often similar to that of the females in that the median apodeme is present in a number of Erirhinine genera as in many Orthoceri weevils. *Tergite 9*: supposedly absent as in all Curculionidae. A detailed description of *H. natans* is provided by Kolbe (1911; in German).

## 2.3 Brief biology of the agent

Female *H. natans* lay eggs singly near the apical tip of the plant stem; oviposition typically occurs on the first division of foliage from the petiole. Eggs have been occasionally also observed oviposited closer to the tips of the foliage, in the petiole itself and once in the floating leaf associated with a flower. Eggs are usually oviposited in a small divot in the plant tissue so that they are around 50% exposed. But this divot can be shallower, so that the egg is more exposed with a thin layer of plant tissue covering the egg, or deeper (Figure 6A). Eggs are shaped like elongate capsules and are creamy in colour, with a length of 855.34  $\pm$  50.17  $\mu$ m (mean  $\pm$  SD) and a width of 292.24 ± 25.51µm (mean ± SD). Eggs hatch about 6 days after oviposition (Kumaran et al. unpublished).



**Figure 6 Life stages of** *Hydrotimetes natans* **(A) Egg laid singly on leaf, (B) first instar larvae, (C) mature larva (D) pupae, (E) mature pupae with adult ready for eclosion, and (F) Adult (Source: CSIRO)** 

The larval phase of *H. natans* lasts for 25-27 days. Larval head capsule widths have three distinct sizes, indicating three larval instar stages. Larvae have an average head capsule width of 190.31 ± 15.68µm (mean±SD), 293.52 ± 16.07µm and 448.98 ± 25.00µm for first, second and third instar, respectively (Kumaran et al. unpublished). First instar larvae are usually found feeding in tunnels in the foliage and petioles of the plant but eventually transition into the main stem. First instar larvae are found as early as 7 days and as late as 24 days after plants are exposed to adults (although only one first instar larva has been found as late as day 24 in our observations). Larvae have a translucent body immediately after hatching from the egg (Figure 6B), which gradually becomes a creamy white colour as they feed. Second instar larvae can be difficult to distinguish morphologically from late first instar larvae but are only found in the main stem of the plant from day 17 to day 29 of development. Third instar larvae are always found in the main stem of the plant from day 18 to day 32 of development. As they develop, their body transitions from a

creamy white to a yellow or green colour and becomes more scarabaeiform (grub shape) as they near pupation (Figure 6C). Prepupae then exit the stem to pupate at a node near the apical tip of the plant (Figure 5D). Pupae turn into darker colour as adults develop (Figure 6E) and adult emergence ensues (Figure 6F).

Larval feeding commences immediately after egg-hatch. As eggs are partially sunken into the plant tissue at oviposition, larvae perhaps chew through the end of their egg case directly into their first tunnel in the plant tissue. First instar larvae feed through the foliage and petiole where their egg was laid, and tunnel through the main stem as they develop into late instar. Their tunnelling causes heavy damage to the foliage and petiole, which is easily observable under a dissecting microscope. Once larvae move into the main stem of the plant, external observations of tunnelling become difficult until they reach third instar. Necrosis of the main stem can be seen by dissecting the plant. Larvae show no preference for which direction they tunnel in the main stem and numerous entry/exit holes are observed on the stem, suggesting that larvae readily exit their tunnels and begin new ones when necessary. Multiple larvae of different instars have been observed feeding in the same section of stem. The presence of larval exuviae within the tunnels suggests that larval moulting occurs within feeding tunnels. Third instar larvae in their feeding tunnels are externally visible as an elongate creamy patch in the main stem.

The average time from oviposition to pupation in our culture has been  $33.8 \pm 5.8$  days, with a further 14.3 ± 2.7 days on average for pupal developmental before adult eclosion (Kumaran et al. unpublished). The average duration of the lifecycle from oviposition to adult emergence has been  $46.5 \pm 4.4$  days (Figure 7). Larval feeding is the principal mode by which the insect damages the plant. Adult feeding, observable through a dissection microscope, is usually focussed on the petioles at the apical growing tip of the plant. Adults produce a single, deep feeding hole each time they feed as they insert their proboscis into the plant tissue. Necrosis is occasionally observed around these feeding scars, but adults seem to otherwise inflict little damage to the plant. Adults are able to persist for several weeks with little or no feeding, suggesting a slow metabolism. No sexual dimorphism has been observed thus far in adult *H. natans*, although females may be slightly larger than males.



**Figure 7 Lifecycle of** *Hydrotimetes natans* **(Source: CSIRO)**

## 2.4 Native range of the agent

*Hydrotimetes natans* has only been recorded in field surveys from Argentina and Paraguay, but possibly occurs in neighbouring countries in South America.

## 2.5 Related species to the agent and a summary of their host range

The genus *Hydrotimetes* has three known species *H. natans, H. striatus* Hust. and *H. tibialis* Hust. The other two *Hydrotimetes* species were described by Hustache in 1926 (http://insecta.pro/taxonomy/195575), but their host records are unknown. Other genera in the Erirhininae are restricted to aquatic plants or are endophagous in roots, stems, leaves, and fruits (Kuschel 1971). Notable species of Erirhininae include the rice weevil *Lissorhoptrus oryzophilus* Kuschel which infests rice crops, and a range of existing or potential biocontrol agents of aquatic weeds: *Cyrtobagous salviniae* Calder & Sands and *C. singularis* Hust. on Salvinia; *Neochetina eichhorniae* Warner and *N. bruchi* Hust. on water hyacinth and *Bagous hydrillae* O'Brien on hydrilla.

## 2.6 The proposed source of the agent

The colony of *H. natans* in quarantine in Australia that has been used for host-specificity testing was established with adults sourced from Iberá wetlands, Corrientes province, Argentina and from wet grasslands in southern Paraguay. *Cabomba caroliniana* plants with *H. natans* collected were air dried in Berlese funnels to extract adult *H. natans* and the adults were hand carried into quarantine in Australia. The imported adult *H. natans* were identified and confirmed by Dr. Rolf Oberprieler (specialist on Coleoptera: Curculionidae, Australian National Insect Collection).

Voucher specimens have been deposited in the Australian National Insect Collection. Colonies of different genetic material will continue to be maintained until permission to release the insect is granted. Additional importations of *H. natans* from S. America may be required to incorporate more genetic diversity in the colony to overcome possible genetic bottlenecks.

Guillermo Cabrera Walsh, Director, FuEDEI is our key South American collaborator for the cabomba biological control project.

## 2.7 Possible interactions with existing biological control programs (of same or related targets, and other targets)

No biocontrol agent has been released in Australia for the biocontrol of *C. caroliniana.*

## 2.8 The agent's potential for control of the target

*Hydrotimetes natans* larvae cause severe damage in the form of necrosis of stem tissue around larval tunnels (Figure 8). Under intense feeding, stems become disintegrated and detached from plants; these decaying stem fragments are not viable, thereby reducing the overall growth and biomass of the plant. This can have a significant effect on the reproduction and spread, as *C. caroliniana* reproduce through healthy stems detaching/fragmenting from the parent plant. Adult feeding is usually concentrated on the petioles at the apical growing tip of the plant. Localised necrosis was occasionally observed around these feeding scars, but otherwise adult feeding does not appear to have a major impact on the plant.



**Figure 8** *Hydrotimetes natans* **larval damage on** *Cabomba caroliniana* **(A) tunnelling of leaves by early instar larvae, (B) larva inside the stem tunnelling, (C) late instar larva and tunnelling damage, (D & E) tunnelling visible through naked eye, (F) tissue decay along the tunnelling (source for all photos: CSIRO)**

A climate matching model was developed using the Composite Match Index in CLIMEX (Kriticos et al. 2015) to predict regions in Australia that most closely match temperatures in the Iberá wetlands of Argentina where *H. natans* is abundant and was collected to establish our colony. The model suggests that regions in south-east QLD and northern NSW that have major infestations of *C. caroliniana* should be highly suitable for *H. natans* (Figure 9). It must be noted however, that the model was developed using only the ambient temperature index in the native range and that other factors could affect establishment of *H. natans* after release. Water temperature, acidity and light penetrability are not taken into account in the model predictions (D Kriticos pers. comm.) and these characteristics of water bodies are believed to affect the abundance of *H. natans* in the native range.



**Figure 9 (A) Regions of Australia that match the climate of the area in the native range (Iberá wetlands, Argentina) where** *H. natans* **was sourced to establish the quarantine colony according to the climate matching model developed. The highest climate matches are represented by red squares, and the moderate and lowest matches are represented by orange and yellow respectively (CMI – Composite Match Index); (B) A red circle is centred on the largest known population of** *C. caroliniana* **in South America (Iberá wetlands in north-eastern Argentina, and the wet grasslands of southern Paraguay) where** *H. natans* **was sourced and the prediction model was developed based on the climate in these regions.** 

Biocontrol of aquatic plants using specialist weevils as biological control agents, has a long legacy of success in weed biological control. Examples include significant control of Salvinia by *C. salviniae*, and control of water hyacinth by *N. bruchi* and *N. eichorniae* in Australia*.* For example, after the first release of *C. salviniae* on an infestation of Salvinia at Lake Moondarra, Mt Isa, Qld in June 1980, the weevils caused severe damage and significantly reduced the weed population within 15 months.

## 2.9 Details of the quarantine facility and methods of containment

Imported *H. natans* are being held within the Approved Arrangement Q2275, BC 5.3 and BIC 7.3, situated at the Ecosciences Precinct, Dutton Park, QLD 4102. All the quarantine-based hostspecificity tests were performed in this facility.

Containment and handling of all imported insects, including killing of required specimens, is being done according to the Department of Agriculture, Water and the Environment (DAWE) guidelines for Approved Arrangements. All staff are experienced and strictly follow the Standard Operating Procedures developed for the facility.

## 2.10 When, when and how initial release will be made

#### **2.10.1 Release from quarantine**

A protocol will be developed with entomologists from DAWE, for the removal of *H. natans* from quarantine. This will likely be similar to protocols developed for previous releases of similar aquatic weevils for weed biocontrol. Once approval for release is obtained, adults will be carefully inspected to ensure that no other associated organisms such as parasite or pathogen are taken from the quarantine. All requirements imposed by DAWE on the release permit will be followed. Once removed from quarantine, *H. natans* will be maintained on *C. caroliniana* in non-quarantine glasshouses while mass rearing protocols are being optimized.

#### **2.10.2 Distribution in the field**

*Hydrotimetes natans* will be released initially into Lake MacDonald and Lake Kurwongbah (QLD), where major infestations of *C. caroliniana* occur. It is expected that water asset managers (e.g. SEQ Water) and council groups (Noosa Landcare) will contribute to releases and distribution of *H. natans*. After initial releases in QLD, subsequent releases will be planned in NSW and NT should C. caroliniana infestations there require long term control and providing biocontrol is compatible with any other ongoing management activities already in place at those sites.

## 2.11 Non-target organisms at risk

See section 1.1.4. Close relatives native in Australia.

## 3 Host-specificity testing

### 3.1 Introduction

The host range of *H. natans* was explored by surveying non-target plants in the field in the native range at sites where *C. caroliniana* occurs and undertaking a series of laboratory host-specificity tests on non-target species in Argentina and Australia, including (i) pilot studies of adult feeding on leaf discs or cut sprigs; and (ii) comprehensive trials with whole plants to assess the life cycle development of *H. natans*. A total of 17 plant species from the Cabombaceae (including the target target (*C. caroliniana*)), Nymphaeaceae and Hydatellaceae families were included in hostspecificity tests.

### 3.2 Surveys of plant use under natural conditions in the native range

Field surveys were undertaken in March and April 2018 to find sites with *H. natans*, study its biology and behavior on its host *C. caroliniana,* examine non-target plants for presence *of H. natans* and collect adults (Figure 10). Observations on the presence of *H. natans* and feeding damage on non-target plants were made at Iberá wetlands in Argentina and three other sites in Paraguay (Encarnación, San Ignacio and Pilar). Aquatic species co-occurring with *C. caroliniana* at these sites were: *Egeria najas* Planch. (Hydrocharitaceae)*, N. indica* (L.) Kuntze (Menyanthaceae)*, N. prolifera* (Wiersema) (Nymphaceae), *Salvinia minima* Baker (Salviniaceae) and *Ludwigia grandiflora* (Michx.) Greuter & Burdet) (Onagraceae). Field surveys revealed that *H. natans* was primarily found on *C. caroliniana*, except for a single adult observed on *N. prolifera* on one occasion. However, no feeding damage on *N. prolifera* was noticed, suggesting that it was likely a casual occurrence of the insect on this plant. The other aquatic plants co-occurring with *C. caroliniana* at Iberá wetlands were *N. jamesoniana* Planch. (Nymphaceae)*, Victoria cruziana* Orb. (Nymphaceae), *L. peploides* (Kunth) P.H.Raven (Onagraceae)*, Utricularia* spp*.* (Lentibulariaceae)*, S. adnata* D.Mitch. (Salviniaceae)*, Eichhornia azurea* (Swartz) Kunth (Pontederiaceae)*, E. crassipes, Hydrocotyle ranunculoides* L.f. (Apiaceae) and *Potamogeton* spp. (Potamogetonaceae). Surveys were also made on *Cabomba haynesii*, *C. furcata*, *C. caroliniana flavida*, and *Najas guadalupensis* (Hydrocharitaceae) populations that do not co-occur with *C. caroliniana*. *Hydrotimetes natans* has not been recorded on any of the co-occurring species in the extensive systematic surveys made for natural enemies of *C. caroliniana* since 2004 (see section 1.10 and Cabrera-Walsh et al. 2011).



**Figure 10 Surveys for occurrence of** *Hydrotimetes natans* **on non-target aquatic plants co-occurring in the native range (A&B) Survey in Iberá Wetlands, Corrientes, Argentina, (C&D) Pilar wet grasslands, Paraguay (Source: CSIRO)**

## 3.3 Laboratory host-specificity tests

#### **3.3.1 Test plant list**

The test plant list was developed based on currently accepted phylogenetic information available in the literature (e.g. Löhne et al. 2008; Borsch et al. 2014) and on the most current angiosperm phylogeny (Chase et al. 2016; Puttick et al. 2020). Test plant species have been selected based on their phylogenetic relationship to *C. caroliniana*, according to the centrifugal phylogenetic method (Briese 2003, 2006; Gilbert et al. 2012; Wapshere, 1974). This method is underpinned by the evidence that specialist herbivores are evolutionarily more likely to feed on non-target species closely related to the target weed than those that are more distantly related. Within the phylogenetic/evolutionary framework, selection of representative test species placed an emphasis on native species, species of economic importance and those that are likely to overlap biogeographically with the target weed, where possible (Figure 11). As previously outlined in section 1.1.4, *C. caroliniana* belongs to the family Cabombaceae within the order Nymphaeales, a basal angiosperm order (Puttick et al. 2020). There are only two species in Cabombaceae in Australia; one being *C. caroliniana*, the other, in another genus, is *B. schreberi*. *Brasenia* is a monotypic genus and has a worldwide distribution (Figure 2) and is not endemic to Australia (Ørgaard 1991). The only other families in the Nymphaeales are Nymphaeaceae (the water lilies) and Hydatellaceae (Puttick et al. 2020). Representatives from all three families of Nymphaeales

(Cabombaceae, Nymphaeaceae and Hydatellaceae) present in Australia have been included for risk assessment of *H. natans* (Figure 11, Table 1).



**Figure 11 Molecular phylogeny of Nymphaeales with the taxonomic relationships between** *C. caroliniana* **and other non-target species used in the host-specificity testing of** *Hydrotimetes natans* **shown in boxes. Australian native species are indicated by an asterisk (Stevens 2001 onwards; Chase et al. 2016)**

Table 1 Plant species tested<sup>#</sup> in Australia and Argentina as part of the risk assessment of *Hydrotimetes natans*, a **candidate biological control agent for** *Cabomba caroliniana***. Accessions of** *C. caroliniana* **from Lake MacDonald, QLD and Burringbar Creek, NSW were used as controls in tests**



*#* The plant test list was submitted to Department of Agriculture, Water and the Environment and made available for feedback from stakeholders and the public for three months (from 10/12/2018 to 01/03/2019).

\*Replacement for *Trithuria cookeana*. Sourcing of *T. cookeana* was unsuccessful despite several field trips made. Hence, this species was replaced with *T. austinensis* and *T. fitzgeraldii.*

\$ *Nymphaea nouchali* was used in the preliminary cut leaf disc study; this is an additional species tested and was not proposed in the original test plant list submitted to DAWE.

#### **3.3.2 Source of Cabomba caroliniana and other test plants**

*Cabomba caroliniana* plants were propagated from stem cuttings (Lake MacDonald, QLD; - 26.402859, 152.947240 and -26.402848, 152.947338; Burringbar Creek, NSW -28.439028, 153.489248) by planting four to six node apical segments into a 1:1 mixture of ADA Nature Aquarium Aquasoil Amazonia (Aqua Design Amano Co. Ltd. Niigata, Japan) and fine, white, washed sand (Richgrow Garden Products, Jandacot, Western Australia). The segments were grown in 160L aquaria at 23°C ± 2°C under T5 fluorescent aquarium plant lights with an 8 hour photoperiod for 4

to 6 weeks or until ready for use in host-specificity testing. Each aquarium was filled with reverse osmosis water and 4g KH buffer (Seachem Laboratories, Madison, GA., USA), 12.5mL of Rexolin APN (Yara Australia Pty. Ltd., McMahons Point, NSW) into which 40mL of a solution containing ammonium sulphate (6.1g/L), potassium nitrate (1.0g/L), magnesium sulphate (1.0g/L) and monopotassium phosphate (0.1g/L) were added. The water was maintained at a pH of 6.2  $\pm$  0.2 using a computer-controlled CO<sub>2</sub> injection system (Aquatronica S.R.L., Reggio Emilia, Italy; Dohse Aquaristik GmbH & Co. KG, Gelsdorf, Germany).

The various non-target plant species tested were either grown from corms supplied by commercial *Nymphaea* growers (*N. immutabilis* S.W.L. Jacobs*, N. pubescens* Willd.*, N. violacea* Lehm.) or collected from the field from natural waterbodies (*B. schreberi, N. alba, N. caerulea, N. gigantea*  Hook.*, N. mexicana, Trithuria spp.,*), under relevant state and territory permits for collection (Appendix II).

All *Nymphaea* species were maintained in 300L fibreglass tanks filled with tap water treated to remove chlorine and amines using Seachem Prime (Seachem Laboratories, Madison, GA, USA) and maintained at a temperature of 29°C ± 5°C in rooftop glasshouses under ambient light conditions. *Brasenia schreberi* was maintained in 160L aquaria at 23°C and pH 6.2 as described for *C. caroliniana* above. All plants were transferred to the quarantine facility for host-specificity testing.

#### **3.3.3 Pilot studies: tests of adult** *Hydrotimetes natans* **feeding on cut leaf discs or on excised whole leaves**

Trials using a leaf-disc or detached plant parts are extremely conservative by design; these trials were setup to gather preliminary information on whether cut plant material of non-target species would be fed upon by the weevil, and to observe and characterise feeding behaviour of adults.

AUSTRALIA: A preliminary adult feeding trial was performed in the quarantine laboratory using cut leaf discs or excised plant material. It included *C. caroliniana* and the following non-target native and introduced species: *B. schreberi*, *N. caerulea, N. gigantea, N. mexicana* and *N. nouchali* Burm.f*.* Leaf discs were used for most species, except for *N. nouchali* and *C. caroliniana* where each replicate consisted of a whole leaf and small sprig, respectively due to the small size of the leaves of the former and the leaf structure of the latter. Each leaf disc or a whole leaf or a sprig of *C. caroliniana* was kept in a round takeaway container (700 ml) with slightly acidic water (pH 6.5) and a pair of *H. natans* was released into each container. After 10 days of the trial, leaves were observed under microscope for feeding lesions and tissue damage. Six replicates per species were studied in this manner.

Feeding lesions caused by *H. natans* adults were observed on *C. caroliniana, B. schreberi, N. caerulea, N. nouchali* and *N. gigantea*, but not on *N. mexicana*. Number of feeding scars (mean ± 1SE) recorded on *C. caroliniana, B. schreberi, N. caerulea, N. nouchali* and *N. gigantea* were 4.50±1.34, 8.00±3.61, 1.00±1.00, 4.80±1.32 and 2.60±1.47 respectively. Though the average number of adult feeding scars recorded on *C. caroliniana* were fewer than on *B. schreberi* and *N. nouchali* these means were not statistically different (F<sub>5, 24</sub> = 2.04; p = 0.109). Furthermore, our observations suggested that *H. natans* feeding on these non-target species was exploratory. As noted above, the use of cut foliage to test for feeding preference is extremely conservative. In addition, cut *C. caroliniana* sprigs are not ideal tissues for adult feeding as adult *H. natans* tend to feed on the petioles at the apical growing tip, a fact which was evident on subsequent trials that used live growing plants (see section 3.3.4).

ARGENTINA: Similar trials to that mentioned above were performed with *C. caroliniana* var*. caroliniana, C. caroliniana* var*. flavida, N. prolifera, N. caerulea* and *Victoria cruziana* Orb. in the laboratory in Argentina (Figure 12). Superficial feeding by adults, and stem tunnelling by larvae, as well as larval survival, were recorded. Larvae did not feed on any of the Nymphaeaceae offered to them. All the larvae died within four or five days on non-target plants. On the other hand, a significant proportion of the larvae (36%) re-entered fresh stems of *C. caroliniana* var. *caroliniana* and completed larval development. Larvae did not complete development on *C. caroliniana* var. *flavida*.

As for adult feeding, mortality was not recorded because we know from previous attempts of laboratory rearing of weevils, that adults can live without food for many weeks. Several feeding lesions by *H. natans* adults were observed on *N. prolifera* and one lesion on *V. cruziana.* In contrast, adult feeding damage was extensive on both *C. caroliniana* varieties.



**Figure 12 Test of larval development of Hydrotimetes natans leaf discs or cut plant material in the laboratory in Argentina. (A)** *Nymphaea prolifera* **leaf discs, (B & C)** *Victoria cruziana* **leaf discs and detached whole leaf of** *V. cruziana* **, (D)** *Cabomba caroliniana* **sprig in a container: larvae can be seen curled around the stems, and (E) A** *H. natans* **larva can be observed in the process of entering a** *C. caroliniana* **stem (Source: FuEDEI)**

#### **3.3.4 Comprehensive tests on whole plants**

#### **Rationale and methods**

A life history-based host-specificity testing approach (i.e. making only those observations that can be recorded without destructive sampling) was adopted for *H. natans* for the following reasons. An adult *H. natans* has a largely aquatic life only surfacing for brief periods of time. It feeds by inserting the proboscis into plant tissue and leaves a small hole, visible only under the microscope. This is also true for oviposition sites and such sites are difficult to locate without staining the plants with food dye (to detect eggs) due to their small size and location (semi-embedded into a plant tissue). Only the stem tunnelling and entry and exit holes of the third instar larvae can be

observed without destructive sampling. Observation of damage caused by the first and second instar larvae requires removal of plant parts and dissection and examination under a microscope.

All non-target test plants included in tests, except for the Hydatellaceae, are submerged aquatic species with only the terminal leaves reaching the water surface. *Cabomba caroliniana* and the non-target plants are all fragile and can be damaged when removed frequently from the water for examination. Such removal can have significant negative impacts on the structural integrity and overall health and condition of the plant, which in turn can affect any larval development. Furthermore, it is not easy to distinguish between damage caused by handling plants during such removal and those caused by *H. natans* feeding and development. Consequently, to overcome such difficulties, host-specificity testing with *H. natans* on whole plants focussed on documenting the likelihood of lifecycle completion on non-target and target plants. Observations on adult and larval feeding and larval development were left to the end of each of the tests when dissections of each plant were undertaken. However, observations for oviposition were made on plants *in situ* during the tests, with extreme care to ensure plants were not damaged (detailed below), because this data could not be collected otherwise and was important to fully assess the host range of *H. natans*. All target and non-target plants used in the whole plant studies were with apical tips/petioles and stems (except for *Trithuria* spp. which has a moss-like/grass-like growth form) to support potential adult and larval feeding and development.

All comprehensive host-specificity tests were conducted within the quarantine facility in a temperature and humidity-controlled glasshouse maintained at a temperature of 27°C ± 4°C with a relative humidity of approximately 40%. All test plant species were subjected to no-choice tests in which potted plants were introduced to either a 68L food grade polypropylene crate or 8L food grade container filled with reverse osmosis water which was treated as described above (in section 3.3.2). Number of plants used in these tests was dependent on size of the plant species and the experimental arenas used; a single plant of *Nymphaea* species but two control plants (*C. caroliniana*) were maintained in 68L crate to match with the larger size of *Nymphaea* species (Figure 13A-D). For tests with *Trithuria* species, a single plant was used for both test plants and the control in 8L container (Figure 13E).

In addition to no-choice trials, choice and continuation trials were also setup with *B. schreberi* (only species on which *H. natans* development was recorded in no-choice trials were progressed to choice and continuation trials). For choice trials, potted plants of both *C. caroliniana* and *B. schreberi* were maintained in a single polypropylene crate (Figure 13F); this experimental set-up allowed *H. natans* to choose between plant species for feeding, oviposition and lifecycle completion. For continuation trials, two potted *B. schreberi* plants were maintained in a polypropylene crate, and additional plants were provided as required to ensure that host plant availability was not a limiting factor for life-cycle completion. This continuation trial was run for ~150 days to determine the ability of *B. schreberi* to sustain a population of the weevil; this duration corresponds to the developmental time of three generations (~42 days per generation) of *H. natans* on *C. caroliniana*.



**Figure 13 Host testing setup (A, B, C & D) Large food grade polypropylene crate with live plants in quarantine glasshouse, (E) Host testing setup of** *Trithuria* **sp. and (F) Setup of choice trials with** *Cabomba caroliniana* **and** *Brasenia schreberi* **(Source: CSIRO)**

Each non-target plant species was tested in six replicates for no-choice tests, except for *T. fitzgeraldii* D.D. Sokoloff et al. (four replicates) and *T. austinensis* D.D. Sokoloff et al. (three replicates) because of difficulties in growing these species (Table 2). For *B. schreberi*, five replicate choice trials and four replicate continuation trials were also setup. In all tests, each replicate of the target (control) or test plant or both in the case of choice trials was randomly allocated to six *H. natans* adults. All adults used in the tests were from a *H. natans* colony maintained on *C. caroliniana*. Since *H. natans* is not sexually dimorphic (determining sex is possible only through dissection of reproductive organs), adults pairing (*in copula*) with each other were used for testing, and equal number of males and females were assumed. Adults were introduced by having them in

a Petri dish and gently submerging the Petri dish in the centre of testing arenas. Upon release, adults were found swimming down towards the bottom of testing arenas both in control and test plant replicates. Further observations showed adults holding on to plant parts and on to the rim of the plant pots both in control and test plant replicates.



**Table 2 Host-specificity testing of** *Hydrotimetes natans:* **number of replications tested in each type of trials for each non-target species from different families**

Observations on oviposition were made three times a week for the duration of each replicate in each trial. If, during the course of the testing, any plant material had to be removed due to decay, it was examined under microscope to check for presence of oviposition marks, eggs, larvae and feeding, to determine whether the cause of the decay was damage from insect activity and observations were recorded. Each trial of host-specificity testing ran for greater than 42 days to correspond with the duration of life cycle of *H. natans* on *C. caroliniana* and multiple trials were run over a three-year period to complete testing of all plant species*.* A control containing *C. caroliniana* was used in each trial of host-specificity testing. At the end of the testing period, each replicate was broken down and all plant material was microscopically examined (Figure 14). Observations on (1) presence of adult feeding and number of adult feeding scars, (2) presence of oviposition, (3) presence of larval feeding, (4) number of pupae formed and (5) number of adults emerged were recorded. This life history-based host testing has allowed us to assess the risk to non-target species from feeding through to lifecycle completion.



**Figure 14 Observations made on the control** *Cabomba caroliniana* **at the end of each trial. (A) Adults recovered, (B) larvae exiting at the base of a plant, (C) Visible larval tunnelling on a stem and (D) live pupae (Source: CSIRO)**

**Statistical analyses and interpretation**

For continuous data (i.e. number of adult feeding scars) a one-way ANOVA was used with test species as a fixed effect (Zar 1999). For binomial data (i.e. presence/absence of oviposition, larval feeding and development, pupation and lifecycle completion) a logistic regression analysis was used to calculate the likelihood of *H. natans* ovipositing or completing its lifecycle on the nontarget test plants (Agresti 2018). For the logistic regression model, *C. caroliniana* was denoted as the reference level, and comparison of non-target species was made against it. In the logistic regression model, the model estimates the log odds of *H. natans* to lay eggs, develop and complete its lifecycle on non-target species. Negative log-odds ('estimate' in the output; See Appendix III) indicates the likelihood of oviposition is lower on the non-target species, while a positive value indicates it is higher, than on *C. caroliniana*.

For the choice trials with *C. caroliniana* and *B. schreberi*, the data were subjected to a binomial test (Zar 1999) to calculate the observed proportion of successful oviposition, larval development, pupation and lifecycle completion on *B. schreberi* (against hypothesised probability of success of 100%). The number of pupae recorded from *C. caroliniana* and *B. schreberi* was subjected to a Welch's t-test (Zar 1999).

All analyses were performed in R3.6.2 (R Core Team 2019) via the RStudio interface (v 1.2.5033) (R Studio Team 2019) using the packages brglm2 (logistic regression; Kosmidis 2019) and ggplot2 (for graphs; Wickham, 2016). For ANOVA, binomial tests and Welch's t-test, R's base functions were used (R Core Team 2019). The R functions, codes and summary output of all analyses are provided as Appendix III. The raw data of host-specificity tests are presented as Appendix IV.

#### **Results**

#### No-choice trials

#### *Nymphaea* species:

Feeding by *H. natans* adults was observed on some of the replicates of *N. gigantea, N. immutabilis, N. pubescens* and *N. violacea* in the no-choice trials. However, the feeding intensity on *C. caroliniana* was far greater compared to that observed on these *Nymphaea* species (F<sub>12,61</sub> = 5.724, p < 0.0001); 217.14 ± 78.05 (mean ± SE) feeding scars on *C. caroliniana* compared to 1.5 ± 0.85 to 16.83 ± 7.70 on the aforementioned *Nymphaea* species (Figure 15). Oviposition by *H. natans* was not observed on any of the replicates of these *Nymphaea* species (Table 5). Neither adult feeding nor oviposition were observed on the other *Nymphaea* species tested: *N. caerulea, N. alba* and *N. mexicana* (Figure 15). The logistic regression analyses revealed that *H. natans* is far less likely to lay eggs and complete its lifecycle on *Nymphaea* species than on *C. caroliniana* (Appendix V).

#### *Trithuria* species:

No feeding and oviposition, and thus no development, by *H. natans* was observed on any of the *Trithuria* species tested (*T. laterna* D.A. Cooke*, T. austinensis, T. submersa* and *T. fitzgeraldii*) under no-choice conditions (Figure 15, Table 3). The logistic regression models fitted for oviposition, larval feeding, larval development, pupation and lifecycle completion suggested that *H. natans* is far less likely to lay eggs and complete its lifecycle on *Trithuria* species than on *C. caroliniana* (Appendix V).



**Figure 15** *Hydrotimetes natans* **adult feeding damage on** *Cabomba caroliniana* **(CC) compared with that on nontarget plant species in no-choice trials (BS –** *Brasenia schreberi***, NAL –** *Nymphaea alba***, NC –** *N. caerulea***, NG – N.**  *gigantea***, NI –** *N. immutabilis***, NM –** *N. mexicana***, NP –** *N. pubescens***, NV –** *N. violacea***, TA –** *Trithuria austinensis***, TF –** *T. fitzgeraldii***, TL –** *T. lanterna***, TS –** *T. submersa***). The error bar on each column corresponds to the standard error**

#### *Brasenia schreberi:*

Adult feeding, oviposition and partial development of *H. natans* were observed on *B. schreberi* in no-choice trials. The intensity of adult feeding was significantly lower on *B. schreberi* than on *C. caroliniana* (Figure 15). Oviposition was observed on four of the six replicates tested. Larval development and feeding were observed in three replicates, but development of larvae through to pupation and lifecycle completion was only recorded on one of these replicates (Table 3). The logistic regression model fitted for oviposition and larval feeding suggested that the odds of *H. natans* laying eggs, and larval feeding are likely. However, the log odds for larval development, pupation and lifecycle completion indicates a significantly lower likelihood for pupation and lifecycle completion by *H. natans* on *B. schreberi* relative to *C. caroliniana* (Appendix V).

**Table 3 Number of replicates in which oviposition, larval development, pupation and lifecycle completion by**  *Hydrotimetes natans* **were recorded on** *Cabomba caroliniana* **(CC) and non-target species in no-choice trails (BS –**  *Brasenia schreberi***, NAL –** *Nymphaea alba***, NC –** *N. caerulea***, NG – N.** *gigantea***, NI –** *N. immutabilis***, NM –** *N. mexicana***, NP –** *N. pubescens***, NV –** *N. violacea***, TA –** *Trithuria austinensis***, TF –** *T. fitzgeraldii***, TL –** *T. lanterna***, TS –**  *T. submersa***)** 



#### Choice trials:

Because of oviposition and partial development of *H. natans on B. schreberi* in no-choice trials, choice trials were setup to determine if the weevil shows preference towards this species over *C. caroliniana*. While oviposition and larval development were recorded on *B. schreberi* in two replicates, *H. natans* did not complete its lifecycle on this species in choice trials.



**Figure 16 Number of replicate choice trials performed with** *Cabomba caroliniana* **(CC) and** *Brasenia schreberi* **(BS) in which oviposition, larval development, pupation and lifecycle completion by** *Hydrotimetes natans* **was recorded**

The results of a binomial test showed that the probability of lifecycle completion by *H. natans* on *B. schreberi* is 0% compared to 100% in *C. caroliniana* (Table 4 & Figure 16). The decrease in the probability from oviposition (40%) through to pupation (20%), and ultimately lifecycle completion (0%) in *B. schreberi* supported the results obtained in no-choice trials. The number of pupae recorded from *B. schreberi* was 0.40±0.40 (mean±SE), which is significantly lower (t = -3.9323; df = 4.22; p <0.05) than that recorded from *C. caroliniana* (10.0±2.41).



**Table 4 One proportion z-test comparing proportion of successful oviposition, larval development, pupation and lifecycle completion by** *Hydrotimetes natans* **between** *Cabomba caroliniana* **and** *Brasenia schreberi*

#### Continuation trials:

Among the four replicates of *B. schreberi*, oviposition was observed on three replicates and larval development and pupation was observed on one replicate; a single pupa was recorded on this replicate and an adult emerged from this pupa. The results of the continuation trials indicated that *B. schreberi* could not sustain multiple generations of *H. natans,* and further strengthened the inference from the no-choice and choice trials.

### 3.4 Discussion

Insect herbivores are evolutionarily more likely to feed on species that are phylogenetically closely related to host plants than those that are more distantly related. Most insect herbivore species tend to specialise and feed on one or a small number of plant species belonging to the same genus, subfamily or family (Jaenike 1990). The centrifugal phylogenetic method is therefore an effective method to explore the host range limit of a species to determine the extent of its host specialisation; its use in classical biological control therefore has a long and strong history (Wapshere 1974; Briese 2003; 2006; Gilbert et al. 2012).

We selected representative non-target plant species of increasing phylogenetically distance to the target weed *C. caroliniana*, to assess the risks that the introduction of the cabomba weevil *H. natans* into Australia would pose to other species. Our studies included species within the Cabombaceae, a family to which the target weed and only one other species, *B. schreberi*, belong in Australia. Since the Cabombaceae family is not well-represented in Australia, the risk assessment was extended to representative species from the two other families in order Nymphaeales, *viz*. the Nymphaeaceae (*Nymphaea* spp.) and Hydatellaceae (*Trithuria* spp.) that are known to occur in Australia.

#### Risk to *Nymphaea* species

The results of our no-choice studies showed minor feeding by *H. natans* adults only on some replicates of four of the seven Nymphaea species tested: *N. gigantea, N. immutabilis, N. pubescens and N. violacea.* This feeding, however, was negligible compared to that observed on *C. caroliniana* and can be only interpreted as exploratory feeding considering the low number of feeding scars and the fact that no oviposition was observed on any of these *Nymphaea* species. In contrast, adult feeding, oviposition and lifecycle completion were observed in all replicates of *C. caroliniana*. Based on these findings, we do not expect ornamental or native *Nymphaea* species to be at risk of attack from *H. natans* in the field.

#### Risk to *Trithuria* species

There was no adult feeding damage nor oviposition on any of the *Trithuria* species tested. *Trithuria* species perhaps lack fundamental feeding and/or oviposition cues necessary to elicit feeding and oviposition by *H. natans*. In addition, *Trithuria* differs greatly in its morphology and growth habits from *C. caroliniana* and it was therefore not surprising that it failed to support oviposition by *H. natans* despite its phylogenetic proximity to the target weed. The leaves of *Trithuria* which are non-tubular sheaths may not be ideal for oviposition and early instar larval development as *H. natans* insert eggs inside the plant tissue, and the early instar larvae tunnel through the leaves of its host plant *C. caroliniana*. Likewise, second and third instar larvae feed inside the stem and larval development lasts for approx. 30 days in *C. caroliniana* which is unlikely to be supported by the short and slender stems of *Trithuria*. The difference in the growth habits of the two species, *Tithuria* being partially submerged while *C. caroliniana* is completely submerged, is also a likely influence on their relative suitability for *H. natans* which is adapted to fully submerged conditions.

#### Risk to *Brasenia schreberi*

Despite oviposition and larval development in a few replicates in both no-choice and choice trials, lifecycle completion by *H. natans* on *B. schreberi* was observed in only one replicate in no-choice trials. This indicated that the possibility of development and lifecycle completion by *H. natans* on *B. schreberi* was significantly lower than on *C. caroliniana*, which supported lifecycle completion in all replicates tested in both choice and no-choice trials. These results suggest *B. schreberi* is physiologically unsuitable for *H. natans* development despite the elicitation of oviposition in some occasions. The fact that *B. schreberi* did not sustain a population of *H. natans* in continuation trials is further evidence that it is an unsuitable host, especially considering that a healthy and productive colony of *H. natans* was maintained on *C. caroliniana*, which has yielded five generations over an eight-month period in our colony. These results suggest that *H. natans* poses negligible risks to *B. schreberi*.

There are additional factors that further lower the risks to *B. schreberi* from *H. natans*. Abundant populations of *B. schreberi* tend to be restricted to higher latitudes (Lloyd & Kershaw 1997), but the bioclimatic models of *H. natans* predict a more subtropical and tropical potential distribution (Figure 9). Therefore, only negligible risks are predicted to the populations of *B. schreberi* that tend to be more abundant in temperate freshwater bodies in south-eastern Australia. *Brasenia schreberi* is of minor importance as an aquarium or ornamental species in Australia (Tropical Plants Database 2020), and hence the risk to commercial aquarium or ornamentals trade is unlikely to occur. Finally, in a global context, *B. schreberi* is not endemic to Australia and has a native range spanning from the Americas through to the Old World. Together, this suggests that the negligible risks posed by *H. natans* are highly unlikely to have significant negative impact on the abundance of *B. schreberi* in natural settings or on its trade by the aquarium or ornamental sectors.

The high level of host-specificity of *H. natans* is not surprising given its phylogenetic position. Weevils in Erirhininae subfamily (marsh weevils) are highly host specific. Notable examples include the biological control agents for water hyacinths, *N. eichhorniae* and *N. bruchiae*, for watermilfoil, *Euhrychiopsis lecontei,* and Salvinia, *C. salviniae*, and monophagous pest species such as the rice water weevils *Lissorhoptrus* spp. and *Afroryzophilus* spp. The biological control agents in the Erirhininae have proven to be extremely effective in managing aquatic weeds.

Overall, our results have showed no risks to phylogenetically proximate species in the genera *Nymphaea* and *Trithuria*, and negligible risk to the more closely related *B. schreberi*. *Brasenia schreberi* is the only species within the Cabombaceae and hence the likelihood of risks to other species in the family Cabombaceae is non-existent in an Australian context. Collectively, our results indicate that *H. natans* will be a highly specific biological control agent. If approved for release, *H. natans* is likely to contribute to the management of *C. caroliniana*, a significant aquatic weed in Australia.

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## Copies of any references referred to in the application

Copies of the references cited in this application are available from the authors upon request.

#### Appendices 5

## Appendix I. Members of the Nymphaeales in Australia, their distribution, origin status and economic importance







Representative species from this list were included in the host testing. While representatives of several native species were included in the testing, some of the species were not included in any of the host testing studies because sourcing of the species was very difficult despite extensive field surveys (e.g. N. macrosperma) or included only in cut leaf discs study because of the difficulty with sourcing and maintenance of the species under laboratory conditions (e.g. N. nouchali).

#### **Species Source Location GPS coordinates\* Cabombaceae** *Cabomba caroliniana* Field Lake Macdonald, Cooroy, Qld -26.402848, 152.947338 Burringbar Creek, Mooball, NSW -28.439028, 153.489248 *Brasenia schreberi* Field Tahbilk Estate Winery billabong, O'Neills Rd, Nagambie (Vic) -36.827172, 145.102728 **Nymphaeaceae** *Nymphaea alba* Field Lake Macdonald, Cooroy, Qld -26.402848, 152.947338 *Nymphaea caerulea* Field Lake Macdonald, Cooroy, Qld -26.402848, 152.947338 Tyagarah, Tee Tree Lake, NSW -28.610473, 153.565126 *Nymphaea gigantea* Field Tyto lakes, Ingham, Qld -18.656649, 146.143374 Keatings Lagoon, Cooktown -15.506610, 145.225205 Caliguel Lagoon, Condamine, Qld -26.982767, 150.112232 Brigalow Creek, Meandarra, Qld -27.325596, 149.875088 Lonesome Creek, Theodore, Qld -24.860487, 150.066691 Mal and Deb Miller, Kime Rd, Midgee via Rockhampton, Qld -23.490852, 150.543915 *Nymphaea immutabilis* Nursery Suncoast water gardens (Paul Lancaster), Fraser Rd, Beerwah -23.385683, 150.645076 *Nymphaea mexicana* Field Caruthers Park, Caruthers Street, Cooroy, Qld -26.406464, 152.912713 Tahbilk Estate Winery billabong, O'Neills Rd, Nagambie (Vic) -36.827172, 145.102728 *Nymphaea pubescens* Nursery Suncoast water gardens (Paul Lancaster), Fraser Rd, Beerwah -26.837696, 152.969780 *Nymphaea violacea* Nursery Suncoast water gardens (Paul Lancaster), Fraser Rd, Beerwah -12.705914, 131.635771 -12.518612, 131.080266 **Hydatellaceae** *Trithuria austinensis* Field Unnamed lake south of Lake Unicup, Frankland River (WA) -34.360765, 116.723190 *Trithuria fitzgeraldii* Field Lake Manaring, Perth Hills (WA) -31.878337, 116.324230 *Trithuria lanterna* Field Bushland off Jenkins Rd, Weddell, NT outside Darwin -12.617556, 131.03141 *Trithuria submersa* Field Waterloo Nature Reserve, Bunbury (WA) -33.329656, 115.758937

## Appendix II. Source of test plant species used in the host-specificity testing of *Hydrotimetes natans* and their locations

**\*** For species sourced from nursery trade, GPS coordinates of possible locations where the plants were collected are provided

Appendix III. R Code and results of statistical analyses

#### **1. Adult feeding scars**

> scars <- read.csv(file.choose(), header = T) > scarsanova <- aov(scars~species, data = scars) > summary(scarsanova) Df Sum Sq Mean Sq F value Pr(>F) species 12 291505 24292 5.724 1.72e-06 \*\*\* Residuals 61 258856 4244 --- Signif. codes: 0 '\*\*\*' 0.001 '\*\*' 0.01 '\*' 0.05 '.' 0.1 ' ' 1 > scarshsd <- TukeyHSD(scarsanova) scarshsd Tukey multiple comparisons of means 95% family-wise confidence level Fit:  $aov(formula = scans ~ species, data = scans)$ *<u>Sspecies</u>*  diff lwr upr p adj CC-BS 2.156429e+02 90.71042 340.57529 0.0000104 NAL-BS -1.500000e+00 -131.14852 128.14852 1.0000000 NC-BS -1.500000e+00 -131.14852 128.14852 1.0000000 NG-BS 2.666667e+00 -126.98185 132.31518 1.0000000 NI-BS 6.500000e+00 -123.14852 136.14852 1.0000000 NM-BS -1.500000e+00 -131.14852 128.14852 1.00000000<br>NP-BS 1.5333338+01 -114.31518 144.98185 0.9999999 NP-BS 1.533333e+01 -114.31518 144.98185 0.9999999 NV-BS 5.000000e+00 -124.64852 134.64852 1.0000000 TA-BS -1.500000e+00 -160.28636 157.28636 1.0000000 TF-BS -1.500000e+00 -146.45145 143.45145 1.0000000<br>TL-BS -1.500000e+00 -131.14852 128.14852 1.0000000 -1.500000e+00 -131.14852 128.14852 1.0000000 TS-BS -1.500000e+00 -131.14852 128.14852 1.0000000 NAL-CC -2.171429e+02 -342.07529 -92.21042 0.0000089 NC-CC -2.171429e+02 -342.07529 -92.21042 0.0000089 NG-CC -2.129762e+02 -337.90863 -88.04375 0.0000138 NI-CC -2.091429e+02 -334.07529 -84.21042 0.0000206 NM-CC -2.171429e+02 -342.07529 -92.21042 0.0000089 NP-CC -2.003095e+02 -325.24196 -75.37709 0.0000514 NV-CC -2.106429e+02 -335.57529 -85.71042 0.0000176 TA-CC -2.171429e+02 -372.10247 -62.18324 0.0006432 TF-CC -2.171429e+02 -357.89192 -76.39380 0.0001114 TL-CC -2.171429e+02 -342.07529 -92.21042 0.0000089 TS-CC -2.171429e+02 -342.07529 -92.21042 0.0000089 NC-NAL 6.750156e-14 -129.64852 129.64852 1.0000000 NG-NAL 4.166667e+00 -125.48185 133.81518 1.0000000 NI-NAL 8.000000e+00 -121.64852 137.64852 1.0000000 NM-NAL -9.947598e-14 -129.64852 129.64852 1.0000000 NP-NAL 1.683333e+01 -112.81518 146.48185 0.9999997 NV-NAL 6.500000e+00 -123.14852 136.14852 1.0000000 TA-NAL 1.776357e-14 -158.78636 158.78636 1.0000000 TF-NAL 3.907985e-14 -144.95145 144.95145 1.0000000 TL-NAL 3.552714e-14 -129.64852 129.64852 1.0000000 TS-NAL 0.000000e+00 -129.64852 129.64852 1.0000000 NG-NC 4.166667e+00 -125.48185 133.81518 1.0000000 NI-NC 8.000000e+00 -121.64852 137.64852 1.0000000 NM-NC -1.669775e-13 -129.64852 129.64852 1.0000000 NP-NC 1.683333e+01 -112.81518 146.48185 0.9999997 NV-NC 6.500000e+00 -123.14852 136.14852 1.0000000 TA-NC -4.973799e-14 -158.78636 158.78636 1.0000000 TF-NC -2.842171e-14 -144.95145 144.95145 1.0000000 -3.197442e-14 -129.64852 129.64852 1.0000000 TS-NC -6.750156e-14 -129.64852 129.64852 1.0000000 NI-NG 3.833333e+00 -125.81518 133.48185 1.0000000 NM-NG -4.166667e+00 -133.81518 125.48185 1.0000000 NP-NG 1.266667e+01 -116.98185 142.31518 1.0000000 NV-NG 2.333333e+00 -127.31518 131.98185 1.0000000 TA-NG -4.166667e+00 -162.95302 154.61969 1.0000000<br>TF-NG -4.166667e+00 -149.11812 140.78478 1.0000000 TF-NG -4.166667e+00 -149.11812 140.78478 1.0000000  $-4.166667e+00 -133.81518 125.48185 1.0000000$ 



#### **2. Logistic regression analyses**

#### Oviposition

```
> hnatans$species <- relevel(hnatans$species,"CC") 
> ovilogist \leftarrow brglm(oviposition \sim species, data = hnatans)
> summary(ovilogist)
```
Call:

 $begin(formula = oviposition ~ species, data = hhatans)$ 

Coefficients: Estimate Std. Error z value Pr(>|z|) (Intercept) 2.708 1.561 1.734 0.0829<br>speciesBS -2.120 1.779 -1.192 0.2333 speciesBS -2.120 1.779 -1.192 0.2333<br>speciesNAL -5.273 2.225 -2.370 0.0178 \*<br>speciesNC -5.273 2.225 -2.370 0.0178 \*<br>speciesNG -5.273 2.225 -2.370 0.0178 \* speciesNAL -5.273 2.225 -2.370 0.0178 \* speciesNC -5.273 2.225 -2.370 0.0178 \* speciesNG -5.273 2.225 -2.370 0.0178 \*<br>speciesNI -5.273 2.225 -2.370 0.0178 \* speciesNI -5.273 2.225 -2.370<br>speciesNM -5.273 2.225 -2.370 speciesNM -5.273 2.225 -2.370 0.0178 \*<br>speciesNP -5.273 2.225 -2.370 0.0178 \* speciesNP -5.273 2.225 -2.370 0.0178 \* speciesNV -5.273 2.225 -2.370 0.0178 \* speciesTA -4.654 2.342 -1.987 0.0469 \* speciesTF -4.905 2.284 -2.148<br>speciesTL -5.273 2.225 -2.370<br>speciesTS -5.273 2.225 -2.370 speciesTL -5.273 2.225 -2.370 0.0178 \* speciesTS -5.273 2.225 -2.370 0.0178 \* --- Signif. codes: 0 '\*\*\*' 0.001 '\*\*' 0.01 '\*' 0.05 '.' 0.1 ' ' 1 (Dispersion parameter for binomial family taken to be 1) Null deviance: 41.316 on 73 degrees of freedom on 61 degrees of freedom Penalized deviance: 29.20399 AIC: 44.204

#### Larval feeding

 $>$  larvalfeedlogist <- brglm(larvalfeeding  $\sim$  species, data = hnatans) > summary(larvalfeedlogist) Call: brglm(formula = larvalfeeding  $\sim$  species, data = hnatans) Coefficients: Estimate Std. Error z value Pr(>|z|) (Intercept) 2.708 1.561 1.734 0.0829 . speciesBS -2.708 1.762 -1.537 0.1243<br>speciesNAL -5.273 2.225 -2.370 0.0178 \* speciesNAL -5.273 2.225 -2.370 0.0178 \* speciesNC -5.273 2.225 -2.370 0.0178 \* speciesNG -5.273 2.225<br>speciesNI -5.273 2.225 speciesNI -5.273 2.225 -2.370 0.0178 \* speciesNM -5.273 2.225 -2.370 0.0178 \* speciesNP -5.273 2.225 -2.370 0.0178 \* speciesNV -5.273 2.225 -2.370 0.0178 \* speciesTA -4.654 2.342 -1.987 0.0469<br>speciesTF -4.905 2.284 -2.148 0.0317 speciesTF -4.905 2.284 -2.148 0.0317 \* speciesTL -5.273 2.225 -2.370<br>speciesTS -5.273 2.225 -2.370 speciesTS -5.273 2.225 -2.370 0.0178 \* --- Signif. codes: 0 '\*\*\*' 0.001 '\*\*' 0.01 '\*' 0.05 '.' 0.1 ' ' 1 (Dispersion parameter for binomial family taken to be 1) Null deviance: 38.485 on 73 degrees of freedom Residual deviance: 18.869 on 61 Penalized deviance: 29.78346 AIC: 44.869 Larval development  $>$  larvaldevlogist  $<-$  brglm(larvaldev  $\sim$  species, data = hnatans) > summary(larvaldevlogist)  $Ca11$ :  $b$ rglm(formula = larvaldev  $\sim$  species, data = hnatans) Coefficients: Estimate Std. Error z value Pr(>|z|) (Intercept) 2.708 1.561 1.734 0.0829<br>speciesBS -4.007 1.851 -2.164 0.0304<br>speciesNAL -5.273 2.225 -2.370 0.0178 speciesBS -4.007 1.851 -2.164 0.0304 \* speciesNAL -5.273 2.225 -2.370 0.0178 \* speciesNC -5.273 2.225 -2.370 0.0178 \*<br>speciesNG -5.273 2.225 -2.370 0.0178 \* speciesNG -5.273 2.225 -2.370 0.0178<br>speciesNI -5.273 2.225 -2.370 0.0178 speciesNI -5.273 2.225 -2.370 0.0178 \* speciesNM -5.273 2.225 -2.370 0.0178 \* speciesNP -5.273 2.225 -2.370 0.0178 \* speciesNV -5.273 2.225 -2.370 0.0178 \* speciesTA -4.654 2.342 -1.987 0.0469<br>speciesTF -4.905 2.284 -2.148 0.0317 speciesTF -4.905 2.284 -2.148 0.0317 \* speciesTL -5.273 2.225 -2.370 0.0178 \* speciesTS --- Signif. codes: 0 '\*\*\*' 0.001 '\*\*' 0.01 '\*' 0.05 '.' 0.1 ' ' 1 (Dispersion parameter for binomial family taken to be 1) Null deviance: 32.367 on 73 degrees of freedom Residual deviance: 16.044 Penalized deviance: 27.35351 AIC: 42.044

#### Pupation

```
> pupationlogist <- brglm(pupation \sim species, data = hnatans)
> summary(pupationlogist) 
Call: 
brglm(formula = pupation \sim species, data = hnatans)
Coefficients: 
            Estimate Std. Error z value Pr(>|z|)<br>2.708 1.561 1.734 0.0829
(Intercept) 2.708 1.561 1.734 0.0829<br>speciesBS -4.007 1.851 -2.164 0.0304
speciesBS -4.007 1.851 -2.164 0.0304 * 
speciesNAL -5.273 2.225 -2.370 0.0178 * 
speciesNC -5.273 2.225 -2.370 0.0178<br>speciesNG -5.273 2.225 -2.370 0.0178
speciesNG -5.273 2.225 -2.370 0.0178 * 
speciesNI -5.273 2.225 -2.370 0.0178 * 
speciesNM -5.273 2.225 -2.370 0.0178 * 
speciesNP -5.273 2.225 -2.370 0.0178 * 
speciesNV -5.273 2.225 -2.370 0.0178 *<br>speciesTA -4.654 2.342 -1.987 0.0469 *<br>speciesTF -4.905 2.284 -2.148 0.0317 *<br>speciesTL -5.273 2.225 -2.370 0.0178 *
speciesTA -4.654 2.342 -1.987 0.0469 * 
speciesTF -4.905 2.284 -2.148 0.0317 * 
speciesTL -5.273 2.225 -2.370<br>speciesTS -5.273 2.225 -2.370
speciesTS -5.273 2.225 -2.370 0.0178 * 
--- 
Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 
(Dispersion parameter for binomial family taken to be 1) 
 Null deviance: 32.367 on 73 degrees of freedom 
Residual deviance: 16.044 on 61 degrees of freedom 
Penalized deviance: 27.35351 
AIC: 42.044 
Lifecycle completion 
> lifecyclelogist <- brglm(lifecycle \sim species, data = hnatans)
> summary(lifecyclelogist) 
Call: 
brglm(formula = lifecycle \sim species, data = hnatans)
Coefficients: 
             Estimate Std. Error z value Pr(>|z|) 
(Intercept) 2.708 1.561 1.734 0.0829<br>speciesBS -4.007 1.851 -2.164 0.0304
speciesBS -4.007 1.851 -2.164 0.0304 * 
speciesNAL -5.273 2.225 -2.370 0.0178 * 
speciesNC -5.273 2.225 -2.370 0.0178 * 
speciesNG -5.273 2.225 -2.370 0.0178 * 
speciesNI -5.273 2.225 -2.370 0.0178 * 
speciesNM -5.273 2.225 -2.370 0.0178 * 
speciesNP -5.273 2.225 -2.370 0.0178 * 
speciesNV -5.273 2.225 -2.370 0.0178 * 
speciesTA -4.654 2.342 -1.987 0.0469<br>speciesTF -4.905 2.284 -2.148 0.0317
speciesTF -4.905 2.284 -2.148 0.0317 * 
speciesTL -5.273 2.225 -2.370 0.0178 * 
species = 5.273 2.225 -2.370--- 
Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1 
(Dispersion parameter for binomial family taken to be 1) 
 Null deviance: 32.367 on 73 degrees of freedom 
Residual deviance: 16.044 on 61 degrees of freedom 
Penalized deviance: 27.35351 
AIC: 42.044
```
#### **3. Choice trial – Binomial test (one proportion z-test)**

```
Cabomba – oviposition, larval development, pupation and lifecycle completion 
> binom.test(5, 5, p = 1.0, conf.level = 0.95) 
        Exact binomial test 
data: 5 and 5 
number of successes = 5, number of trials = 5, p-value = TRUE
alternative hypothesis: true probability of success is not equal to 1 
95 percent confidence interval: 
 0.4781762 1.0000000 
sample estimates: 
probability of success 
 1 
Brasenia – oviposition & larval development 
> binom.test(2, 5, p = 1.0, conf. level = 0.95) Exact binomial test 
data: 2 and 5 
number of successes = 2, number of trials = 5, p-value < 2.2e-16alternative hypothesis: true probability of success is not equal to 1 
95 percent confidence interval: 
 0.05274495 0.85336720 
sample estimates: 
probability of success 
                    0.4 
Brasenia – pupation 
> binom.test(1, 5, p = 1.0, conf. level = 0.95) Exact binomial test 
data: 1 and 5 
number of successes = 1, number of trials = 5, p-value < 2.2e-16alternative hypothesis: true probability of success is not equal to 1 
95 percent confidence interval: 
 0.005050763 0.716417936 
sample estimates: 
probability of success 
                    0.2 
Brasenia – lifecycle completion 
> binom.test(0, 5, p = 1.0, conf. level = 0.95) Exact binomial test 
data: 0 and 5 
number of successes = 0, number of trials = 5, p-value < 2.2e-16alternative hypothesis: true probability of success is not equal to 1 
95 percent confidence interval: 
 0.0000000 0.5218238 
sample estimates: 
probability of success 
 0
```
#### **4. Choice test – Number of pupae**

```
> bartlett.test(pupae \sim species, data=choicetestpupa)
        Bartlett test of homogeneity of variances 
data: oviposition by species 
Bartlett's K-squared = 8.0305, df = 1, p-value = 0.0046 
# if p value is more than 0.05, var.equal = TRUE
 > t.test(pupae ~ species, data=choicetestpupa, var.equal=FALSE, conf.level=0.95
\frac{2}{1} Welch Two Sample t-test 
data: oviposition by species 
t = -3.9323, df = 4.2205, p-value = 0.01539 
alternative hypothesis: true difference in means is not equal to 0 
95 percent confidence interval: 
 -16.240752 -2.959248
```

```
sample estimates: 
mean in group BS mean in group CC 
 0.4 10.0
```
### **5. Cut leaf disc study – Number of feeding lesions**

```
> lesions <- read.csv(file.choose(), header = T)
> lesionsanova <- aov(lesions~species, data = lesions) 
> summary(lesionsanova) 
            Df Sum Sq Mean Sq F value Pr(>F) 
species 5 223.3 44.66 2.04 0.109 
Residuals 24 525.5 21.90
```
 $\overline{\phantom{a}}$ 

## Appendix IV. Raw data for comprehensive laboratory host-specificity testing

## **Pilot trials**

Table S1. Number of feeding lesions in cut leaves or sprigs of target and test plants in pilot trials



\* Entire leaves of this species were used given its small size

### **No-choice trials**

Table S2. Number of adult feeding scars



#### Table S3. Presence of oviposition  $(1 = yes; 0 = no)$



Trithuria lanterna				
Trithuria austinensis				
Trithuria submersa				
Trithuria fitzgeraldii				

Table S4. Presence of larval feeding  $(1 = yes; 0 = no)$ 



#### Table S5. Presence of larval development  $(1 = yes; 0 = no)$



#### Table S6. Presence of pupae  $(1 = yes; 0 = no)$



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#### Table S7. Lifecycle completion  $(1 = yes; 0 = no)$



## **Choice trials**

Table S8. Oviposition, larval development, pupation and lifecycle completion in choice trials



Table S9. Number of pupae in choice trials





## **Continuation trials**

Table S10. Oviposition, larval development, pupation and lifecycle completion in continuation trails on *Brasenia schreberi* (1 = yes, 0 = No)



\* Only a single pupa and adult emerged from this replicate.



Appendix V. Summary statistics of logistic regression analyses performed with *Cabomba caroliniana* as a reference species

BS – *Brasenia schreberi*, NAL – *Nymphaea alba*, NC – *N. caerulea*, NG – N. *gigantea*, NI – *N. immutabilis*, NM – *N. mexicana*, NP – *N. pubescens*, NV – *N. violacea*, TA – *Trithuria austinensis*, TF – *T. fitzgeraldii*, TL – *T. lanterna*, TS – *T. submersa*

The 'estimate' is a slope of the logistic regression model and is the log-odds; the *p* value indicates statistical significance of the model comparing the log-odds for a given species relative to the reference species *Cabomba caroliniana*.

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